

Publication status: This preprint has been published elsewhere.

DOI of the published preprint: <https://doi.org/10.1590/2675-2824073.25025>

# Microplastic Hotspots on a Tropical Estuarine-Bay System

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<https://doi.org/10.1590/SciELOPreprints.12903>

Submitted on: 2025-08-06

Posted on: 2025-08-06 (version 1)

(YYYY-MM-DD)

## Microplastic Hotspots on a Tropical Estuarine-Bay System

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### ABSTRACT

The accumulation and export of microplastics are associated with growing negative socio-environmental impacts, which are intensified in coastal and estuarine areas. The Santos Estuarine System is the most populous estuarine region and the largest seaport in Brazil, where the movement and accumulation of microplastics is an integrated effect of diverse hydrodynamic conditions. To identify the vulnerability of pollution by microplastics in extreme conditions of sea surface elevation in this complex estuarine region, the coupling between Lagrangian and Eulerian modelling was employed, an approach that has become popular in the study of ocean movements. The trajectories of microplastics from potential sources were modelled using Opendrift, forced by currents from the Eulerian model Estuarine, Coastal and Ocean Model (ECOM), and waves from Wavewatch III. Particle concentrations and three-dimensional stochastic probability of contact with the continent were evaluated, demonstrating that, regardless of the sea surface elevation condition, microplastics are more exported to the South Brazilian Bight than accumulated on the margins and/or bottom of the system. During higher elevations of the sea surface, there is a greater accumulation of microplastics on the estuary margins, with spring tide promoting faster accumulation. Although the entire coastline of Santos and adjacent beaches are prone to the largest accumulations, portions further to the Northeast of the estuary have not shown the propensity to be contaminated with microplastics released into estuarine channels. The highest concentrations are actively drifting in the system's water column.

**Keywords:** Lagrangian drift, ocean modelling, plastic pollution, Opendrift

### 1 INTRODUCTION

2 Marine pollution is the first of the challenges of the  
3 United Nations Oceans Decade, being relevant to  
4 understanding and mapping sources of pollutants  
5 and contaminants released into the sea, in addition  
6 to their potential impacts on human health and  
7 marine ecosystems (MacLeod et al. 2021). The  
8 accumulation of microplastics – due to widespread

9 use for its production and logistics characteristics  
10 (Lebreton et al. 2017, Sterl et al. 2020) – is notably  
11 riously a form of growing pollution, having already  
12 been detected in all marine environments (Van Sebille  
13 et al. 2012). Plastic degradation occurs due  
14 to environmental conditions, leading to plastic pollution  
15 from macro to microscale (< 5 mm) (Poulain  
16 et al. 2018). Plastic particles on such a scale affect  
17 the biosphere metabolically and structurally  
18 (Kershaw et al. 2019), as well as ecologically, encompassing  
19 impacts on food webs and changes in biogeochemical cycles  
20 (Egger et al. 2020, Malli et al. 2022).

22 The coastal region is where the initial phase of

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23 the processes of variations in the size and proper- 71  
24 ties of plastic debris takes place, due to the high en- 72  
25 ergy of the coastal and estuarine systems combined 73  
26 with the presence of abrasive elements (Kershaw 74  
27 et al. 2019, Hardesty et al. 2017). However, the 75  
28 use of numerical models to track and predict these 76  
29 particles is more often with low spatial resolutions (> 77  
30  $1/12^{\circ}$ ) to the open ocean (Krelling et al. 2017, Kaan- 78  
31 dorp et al. 2023). 79

32 The Santos Estuarine System (SES) ( $24^{\circ}\text{S}$ , 80  
33  $46.3^{\circ}\text{W}$ ), on the South Brazil Bight (SBB), is an 81  
34 important study area, covering 30 km in an east- 82  
35 west direction and 19 km in a north-south direction 83  
36 (Marengo et al. 2017). Due to its interconnected 84  
37 estuarine channels, it has two connections with the 85  
38 ocean and houses the largest maritime port in Brazil 86  
39 (Porto de Santos) (Costa et al. 2020). 87

40 In an estuarine system such as Santos, the main 88  
41 factors controlling plastic drift can be summarised in 89  
42 the tide, abundance of microplastics, rainfall, wind, 90  
43 Stokes drift, and buoyancy. The effect of the ebb 91  
44 tide is to transport particles from the estuary to the 92  
45 ocean, and in flood tide to reintroduce such particles 93  
46 (Malli et al. 2022). Furthermore, the abundance of 94  
47 microplastics is higher in the estuary beds during 95  
48 the neap tide, and in the water column during spring 96  
49 tide (Gorman et al. 2020). 97

50 The increase in flow associated with rainfall is ca- 98  
51 pable of resuspending microplastics in the water col- 99  
52 umn (Piehl et al. 2020), and remobilising particles 100  
53 trapped in vegetation cover (Pinheiro et al. 2021). 101  
54 Concomitantly, the wind affects the suspension and 102  
55 deposition of microplastics (Strand et al. 2021), be- 103  
56 ing able to import or export low-density films and mi- 104  
57 croplastics to the estuary. Furthermore, microplas- 105  
58 tics are redistributed in the water column due to 106  
59 turbulent mixing caused by wind. Wind also causes 107  
60 Stokes drift - due to the displacing of the particle 108  
61 from its final position after a wave period (Dagestad 109  
62 & Röhrs 2019). 110

63 An anomalous situation of meteo-oceanographic 111  
64 conditions in a coastal and estuarine system, with 112  
65 great economic and environmental impact, are the 113  
66 storm surges - a constructive sign of the astronom- 114  
67 ical and meteorological tide with the influence of 115  
68 the local geomorphology (Costa et al. 2020). Storm 116  
69 surges influence the dispersion and accumulation 117  
70 of pollution in the coastal zone (Seiler et al. 2020). 118

Such a direct influence of sea surface elevation on the transport of pollutants becomes more prominent in matters of climate change, and the consequent increase in the frequency of cold fronts, storm surges, as well as the epitome of the general rise in mean sea level (Goldenberg et al. 2001).

Several coastal-scale studies have investigated the transport and accumulation of plastic debris in Brazil. In the Paranaguá estuary, research has tracked the movement of plastic waste (Krelling et al. 2017). In the Santos estuary, seasonal variability in microplastic accumulation has been analysed, highlighting differences between wet and dry periods (Gorman et al. 2020, Balthazar-Silva et al. 2020). Additionally, a method for predicting plastic contributions using remote sensing has been proposed (da Silva Ferreira et al. 2021).

A key approach for studying pollutant dispersion is Lagrangian drift simulation, which enables rapid analysis under various physicochemical conditions (Sterl et al. 2020). However, despite its advantages, this method still presents challenges in accurately identifying when and where pollutants enter and leave different oceanic systems.

In the Santos Estuarine System (SES), sea surface elevation plays a crucial role in debris accumulation near the estuary and along the sandy beaches of Santos Bay (da Silva Ferreira et al. 2021). Additionally, wind-driven current inversions, particularly those associated with the passage of cold fronts from the southwest, further influence plastic transport patterns (Costa et al. 2020). To better understand these processes, this study integrates three numerical models: an Eulerian model for large-scale circulation, a Lagrangian model for particle tracking, and a spectral model for surface wave dynamics.

Thus, to evaluate the transport and accumulation of microplastics in the SES under the influence of different sea surface elevation conditions, this study tested the hypothesis that (i) the dynamics of the SES export microplastics to SBB under normal and extreme conditions (positive and negative) of elevation of the sea surface, and that (ii) under conditions of extreme elevation (lowering) of the sea surface, the system acts to retain a greater (lesser) amount of microplastics. By coupling Lagrangian, Eulerian, and wave models, we sought to identify the hotspots

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119 of microplastic accumulation in Santos Bay and to  
 120 quantify the annual export of microplastics from the  
 121 SES to the SBB.

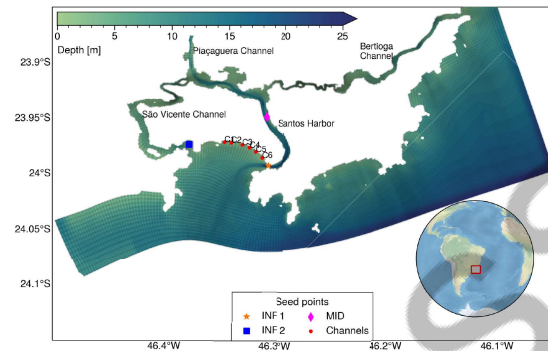
## 122 METHODS

### 123 STUDY AREA

124 The SES is located in the central and widest part  
 125 of the South Brazil Bight (SBB), and is classified as  
 126 partially mixed (Hansen & Rattray 1962). The SES  
 127 is under the influence of high and low-frequency  
 128 physical processes, such as tides, sea surface ele-  
 129 vation, and river discharge, the latter being related  
 130 to longitudinal baroclinic salinity gradient (Costa  
 131 et al. 2020). Composed of four main estuarine chan-  
 132 nels (São Vicente, Piaçaguera, Bertioga and Porto  
 133 de Santos), the SES is of great importance in large-  
 134 scale activities, especially industries in Cubatão and  
 135 the Santos' Harbour (Figure 1). The Santos Bay  
 136 has five artificial channels responsible for leading  
 137 diffuse rainfall from the city into the beaches, which  
 138 are regulated by floodgates. They are kept closed  
 139 during periods of low rainfall to prevent pollution due  
 140 to the irregular discharge of sewage (da Silva Fer-  
 141 reira et al. 2021). However, such gates are opened  
 142 in periods of significant rainfall, making the disper-  
 143 sion and transport of pollutants to Santos Bay an  
 144 important factor.

145 The SES connects to the Inner Continental  
 146 Shelf (ICS), where currents and sea surface height  
 147 (SSH), are controlled by density gradients, tides  
 148 and wind shear stress (Dottori & Castro 2009, Dot-  
 149 tori & Castro 2018). Density gradients are caused  
 150 by the discharge of continental waters which,  
 151 when encountering waters from the shelf, generate  
 152 geostrophic currents parallel to the coast (Costa  
 153 et al. 2020). Wind shear presents the highest vari-  
 154 ability in driving currents and surface elevation. Pre-  
 155 vailing easterly and northeasterly winds, intensified  
 156 in summer, are the main pattern in this region, trans-  
 157 porting surface waters offshore, and reducing the  
 158 SSH (Dottori & Castro 2009). During the passage  
 159 of frontal system events, winds transport waters to-  
 160 wards the coast, contributing to increasing the SSH  
 161 (Stech & Lorenzetti 1992).

162 The M2 tidal harmonic is the most energetically  
 163 significant (Dottori & Castro 2018). However, the



**Figure 1.** Santos Estuarine System bathymetry with its geographic location on the Brazilian coast (red square at the lower right globe), and the locations of seeding locations at the two stages of the experiments by coloured markers. The red circles represent the locations of the drainage channels, labelled with the respective number  $n$  ( $C_n$ ).

164 tides are one order of magnitude lower than the  
 165 perpendicular currents generated by the combina-  
 166 tion of wind shear stress and the density gradient  
 167 (Costa et al. 2020).

### 168 NUMERICAL MODELS

#### 169 *The Eulerian ECOM Model*

170 The current velocity results simulated with the Es-  
 171 tuarine and Coastal Ocean Model (ECOM) serve  
 172 as an Eulerian basis for the OpenDrift Lagrangian  
 173 framework. The ECOM – coastal and estuarine  
 174 three-dimensional version of the Princeton Ocean  
 175 Model (POM) (Blumberg & Mellor 1987) – simulates  
 176 the conservative variables (temperature, pressure  
 177 and salinity), currents and sea level through the so-  
 178 lution of the momentum and continuity equations  
 179 in an Arakawa C-grid in vertical sigma coordinates.  
 180 Analytically, the ECOM employs the hydrostatic and  
 181 the Boussinesq approximations, considering the  
 182 density field perturbations with the vertical momen-  
 183 tum component. Thus, there are prognostic vari-  
 184 ables such as the zonal, meridional and vertical  
 185 components of velocity, temperature, salinity and  
 186 elevation of the free sea surface through the Navier-  
 187 Stokes momentum equations, continuity, and con-  
 188 servation of temperature and salinity, in sigma ver-  
 189 tical coordinates (terrain following) (Blumberg &

Mellor 1987).

The SES curvilinear grid comprises the four main estuarine channels and Santos Bay, as well as the adjacent portion of ICS (Figure 1). It has 322 cells in the parallel direction by 132 cells in the perpendicular direction to the coastline in Santos Bay. Curvilinear grid lines are drawn to create an orthogonal grid with a low aspect ratio and relatively high horizontal resolution within estuarine channels. The average resolution is 144.5 m, ranging from 26.7 m at the São Vicente and Bertioga to 444.5 m at the southwest portion of the domain. Along the Santos Harbour, the horizontal resolution is approximately 100 m (Costa et al. 2020).

The simulations carried out in this study used the operational PreAMar system (Costa et al. 2020), developed to provide daily updated 72-hour forecasts of sea surface elevation and currents in the SES, forced by winds, tides and fluvial discharge (Costa et al. 2020). PreAMar presents its forecast based on seven tidal components (M2, S2, O1, K2, N2 and Q1) with phase distribution along the edge. The initial thermohaline conditions are from climatology (22.3°C, 30.4 PSU), and the thermohaline boundary conditions come from a regional domain over the SBB, as well as the sea and tidal free surface elevation boundary conditions. River discharge is variable through space with 13 freshwater discharge points, also varying through the seasons. Such Eulerian results were validated for the region and period covered by the simulations of the present study. (Costa et al. 2020).

### **The WaveWatchIII Wave Model**

Significant wave height, Stokes drift and average wave period results were generated by the WaveWatch III wave model (ST6 formulation), which is a third-generation model developed at NOAA/NCEP based on Wave Modelling (WAM) (Komen 1994).

The WaveWatchIII equations include refraction and deformation of the wave field due to temporal and spatial variations in mean water depth, and mean current (e.g: tides, storm surges.). Wave propagation is considered linear, and relevant non-linear effects, such as resonant interactions, are included in the source terms (Cavaleri et al. 2007).

The wave simulations consist of nesting two grids (global and South Atlantic grid), which provide the

wave spectrum for an unstructured grid over the continental shelf off the state of São Paulo through its boundary conditions. The global and spatial grid resolution is respectively 2° followed by a 0.5° regional grid covering the São Paulo coast (27°S-22°S, extending to 200m depth), constrained at the edges by the directional wave spectrum generated through the nested simulation. Finally, this São Paulo coast unstructured grid is interpolated to match the high-resolution PreAmar grid within the Santos Estuarine System (SES).

### **The OpenDrift Lagrangian model**

OpenDrift is an open-source framework based on the Python programming language, highly generic and modular for numerical Lagrangian drift modelling (Dagestad et al. 2018), in continuous development, through which recent studies investigated connections of biogeochemical processes in the South Brazil Bight (Mendes et al. 2022), and bacterial dispersion at Cananéia-Iguape Estuarine Complex (Birocchi et al. 2025). This framework is an offline model for computing Lagrangian drift, which is computationally economic when compared to an online model, in which the drift is computed along with the hydrodynamic fields in the Eulerian model.

The OceanDrift module – on which the PlasticDrift used here is based – is the fundamental module for passive particle drifting – particles that quickly adjust their velocities to the velocity field and do not affect the flow properties, thus satisfying the equations of the movement (Prants 2014). This module includes Stokes drift forcings and wind drag, among other easily enabled configurations (Dagestad & Röhrs 2019). In PlasticDrift, particles with sizes between 2 mm and 5 mm, are subjected to three-dimensional transport by currents, surface Stokes drift due to gravity waves, and wind drag. The vertical movement consists of buoyancy and mixing by oceanic turbulence, in which a random scheme of diffusivity in vortices depends on the wind (Strand et al. 2021).

The fourth-order Runge-Kutta scheme was used to integrate over a specific time interval and result in the position of the drifter in each time step, which duration was set as 60 seconds, through bidimensional nearest neighbour interpolation (Dagestad et al. 2018). Such a scheme assumes that a

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284 particle is passive in a three-dimensional flow  
 285 (Edwards et al. 2006). The integrated first-order  
 286 partial differential equation with initial value condi-  
 287 tion (Ličer et al. 2020) is then solved to incorporate  
 288 the Eulerian current, wind drag deviations (Callies  
 289 et al. 2017), and the Stokes drift.

## 290 EXPERIMENTS

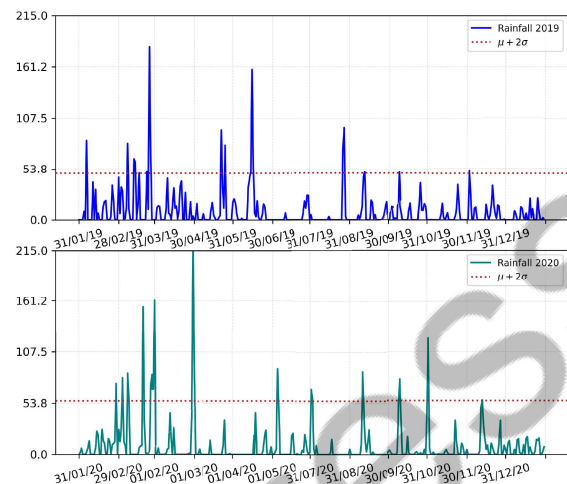
291 To assess the annual export of microplastics from  
 292 the SES to SBB, the dispersion of particles was  
 293 simulated using OpenDrift for two subsequent years  
 294 (2019 and 2020), with inert particles that do not  
 295 interact with the temperature and salinity of the  
 296 environment.

297 We chose 8 seeding locations in the SES domain  
 298 (Figure 1). Six points are located near the artificial  
 299 channels (C1 - C6) and two are arbitrary points  
 300 (INF1 and INF2), chosen based on the surface and  
 301 subsurface sources of solid waste in a Bay, such  
 302 as factories, large urban centres, and junction flows  
 303 (Kaandorp et al. 2021, Balthazar-Silva et al. 2020).  
 304 For each seeding point, neutral buoyancy particles  
 305 were released every hour throughout the whole 2  
 306 years of simulations.

307 Henceforward, beaching refers to the permanent  
 308 retention of particles upon contact with coastal land  
 309 (dry shoreline). Stranding is defined as beaching or  
 310 the seabed contact without resuspension. So, there  
 311 is no longer movement of the particles when they  
 312 contact both the dry cells and the bottom of a wet  
 313 cell

314 The artificial channels are opened during periods  
 315 of intense rainfall to avoid urban flooding; however,  
 316 the information about the opening of these channels  
 317 is not available for public access, so we infer their  
 318 opening based on the prediction regime of interest.  
 319 For that, we used daily accumulated precipitation  
 320 time series provided by the public administration of  
 321 Santos for 2019 and 2020 (Figure 2), defining here  
 322 hypothetical openings on the day and the follow-  
 323 ing day of floodgates for a threshold of 2 standard  
 324 deviations above the average rainfall.

325 High variability in marine litter abundance, includ-  
 326 ing tidal swings, is especially evident in estuarine  
 327 environments (Bettim et al. 2021). Monitoring a fre-  
 328 quency comparable to spring and neap tidal cycles  
 329 provides the necessary information to better under-



**Figure 2.** Rainfall [mm] at Santos City to 2019 (top) and 2020 (bottom). The red horizontal line indicates 2 standard deviations ( $\sigma$ ) from the temporal average ( $\mu$ ).

330 stand the influence of meteorological and oceanographic  
 331 conditions on the amount and type of debris  
 332 on different coastlines (Moreira et al. 2016). There-  
 333 fore, to quantify the vulnerability to contamination  
 334 by microplastics in the SES in events of sea surface  
 335 extreme elevation, the years 2019 and 2020 were  
 336 simulated independently in OpenDrift in periods of  
 337 72 hours, to resolve half of a cycle of neap and  
 338 spring tides.

339 We defined three categories of sea surface elevation  
 340 based on statistical thresholds applied to the sea surface  
 341 height ( $\eta$ ) from the mean ( $\mu$ ) and the standard deviation  
 342 from the mean ( $\sigma$ ): (i) Extreme elevation, when the  
 343 sea surface height exceeded two standard deviations  
 344 above the mean ( $\eta > \mu + 2\sigma$ ); (ii) Extreme decline,  
 345 when it fell below two standard deviations from the  
 346 mean ( $\eta < \mu - 2\sigma$ ); and (iii) Normal elevation,  
 347 when it remained within two standard deviations of  
 348 the mean ( $\mu - 2\sigma \leq \eta \leq \mu + 2\sigma$ ).

349 At each location (INF1, INF2, MID, and artificial  
 350 channels C1–C6; see Figure 1), we released  
 351 42 neutrally buoyant particles per hour. These re-  
 352 leases occurred every 72 hours over the course of  
 353 a year. Particles were seeded at depths between  
 354 0 and 5 m to ensure accurate representation of  
 355 both surface dispersion and subsurface transport  
 356 dynamics, following the methodology of (Hardesty  
 357 et al. 2017).

## ANALYSIS

The particle amount per unit area (km<sup>2</sup>) was calculated separately for 2019 and 2020 to estimate the annual export of microplastics to the continental shelf.

To quantify contamination, probability maps were generated to illustrate the likelihood of particles making contact with land, highlighting the regions within the Santos estuary and bay that are most prone to microplastic accumulation (Saçu et al. 2021).

The three-dimensional contact probability ( $P$ ) is given by:

$$P(x, y, t) = \frac{1}{S} \sum_{\eta=1}^S \sum_{\tau=0}^T \frac{C(x, y, t)_{\eta, \tau}}{N_{\eta, \tau}} \quad (1)$$

where  $C(x, y, t)$  is the number of particles' contacts in a grid cell centred on  $x$  and  $y$  at a given time  $t$ , which varies between 0h and the total time of the simulation.  $N$  is the total number of simulated particles in time  $t$ , and  $S$  is the number of simulations at a given elevation configuration ( $\eta$ ) through period ( $\tau$ ). Naturally,  $P$  ranges from 0 to 1 (Saçu et al. 2021). Through this probability, the temporal variation of the spatial movement of microplastics is evaluated.

The detection of accumulation hotspots was based on identifying regions where the probability of contact exceeded 5% , considering all particles released into the system. Each sea surface elevation condition was analysed separately, and the probability was evaluated at three-times intervals: 24, 48, and 72 hours.

For each 72 hours under a given elevation condition, the contact probability was calculated and then normalised by the total number of 72-hour periods corresponding to that specific elevation condition.

## DISPERSION VALIDATION AND CAVEATS

The main caveat of this work lies in the absence of *in situ* data for validation during the simulated periods. Nevertheless, by employing an Eulerian model that has been previously validated for this region and period, we compared the results of our simulations with the particle dispersal areas reported by Gorman et al (2020). Their study used a simplified

Lagrangian approach, coupled with a coarser Eulerian model configured with a horizontal resolution fixed at 1 km and run for the year 2008.

The seeding points used in the present study were based on those of Gorman et al. (2020), with further adaptations based on Kaandorp et al. (2021). While the particle release locations are not identical, they are conceptually consistent. Given that the Eulerian model employed here has been validated and is capable of capturing the dominant and recurrent physical processes operating during the years 2019 and 2020 (Costa et al. 2020), we argue that the resulting dispersal areas can be reasonably compared with those from the earlier study.

The monthly-averaged dispersal patterns obtained here (presented in the next section in area values and percentual particle amount) are, in both percentual extent and spatial distribution (footprint), qualitatively similar to those presented by Gorman et al. (2020). Specially the seeded particles at INF1, INF2 and the artificial channels that match the beach region at Gorman et al. (2020), that spanned from around 70 km<sup>2</sup> -80 km<sup>2</sup> for INF1 and INF2, and near 20 km<sup>2</sup> for each channel.

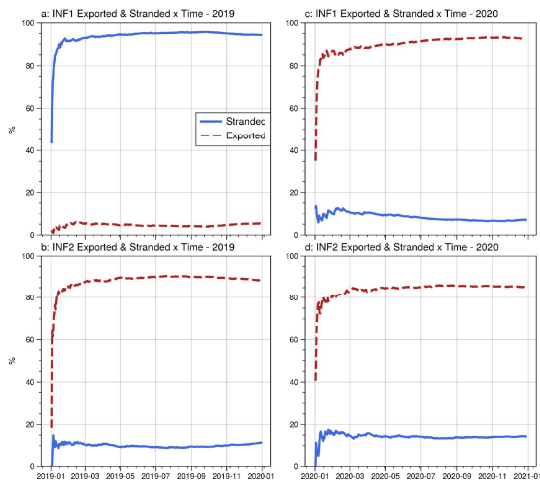
We acknowledge that comparing results from two different modelling studies — rather than directly validating against observational data — is not ideal. However, in the absence of microplastic observational data for the specific years simulated and at matching spatial resolutions, this constitutes a form of indirect validation currently feasible, once Gorman et al. (2020) made *in situ* sampling at the Santos' beaches for 2008. The reader is referred to Gorman et al. (2020) for further details about MP footprint.

## RESULTS

### NATURAL SOURCES OF CONTAMINATION

Figure 3 shows the time evolution of the particles released in INF1 and INF2, for 2019 and 2020, and the time that the particles are stranded at a dry cell (stranding time henceforward).

In 2019, the particle accumulation and exportation presented inverse patterns between INF1 and INF2 (Figure 3). For the INF1 (INF2) releasing point, particles presented an accumulation (exportation)



**Figure 3.** Cumulative export and stranding percentage by time for particles released in 2019 and 2020 in INF1 (a,c) and INF2 (b,d)

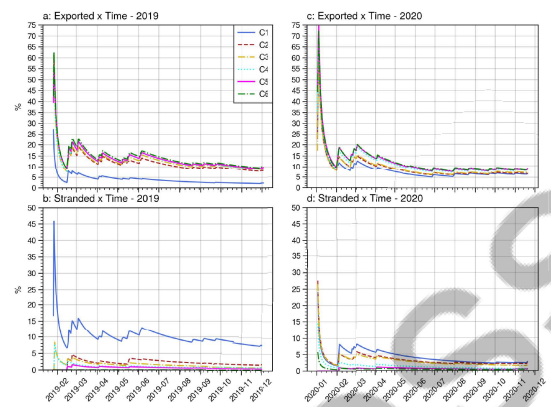
446 within the SES. Particularly for the INF1 location, 447 this pattern affects directly the rate of particle ac- 448 cumulation, resulting in 94.6% particles stranded 449 and only 5.4% exported to the SBB. Particles from 450 INF1 reached an area of 70.37 km<sup>2</sup>, while INF2 451 reached a much larger area of 130.48 km<sup>2</sup>, due to 452 the fast stranding of particles released in INF1, in 453 the channel between Santos and Guarujá.

454 The 2020 scenario exhibited a similar pattern of 455 stranding and exportation between seeding points 456 INF1 and INF2. Throughout the analysis, particles 457 were more frequently exported than stranded within 458 the system (Figure 3c and d).

459 As in 2019, the export-to-stranding ratio influ- 460 enced the dispersion area. However, unlike the 461 previous year, the difference in coverage area be- 462 tween the two seeding points was approximately 463 30%. While INF1 covered an area of 88.65 km<sup>2</sup>, 464 INF2 reached a larger area of 122.78 km<sup>2</sup>. Thus, 465 the exportation rates between the seeding points in 466 2020 were similar: 92.6% for INF1 and 85.3% for 467 INF2.

### 468 ARTIFICIAL SOURCES OF CONTAMINATION

469 Figure 4 illustrates the inverse relationship between 470 particle export from the SES artificial channels and 471 stranding along the coast or bay floor. Higher export 472 rates corresponded to lower stranding rates. The



**Figure 4.** Cumulative percentage export (a,c) by time and stranding (b,d) by time for particles released in the artificial channels of Santos in 2019 and 2020.

473 logarithmic pattern shows periodic spikes coinciding 474 with particle release events, indicating that both 475 export and stranding occur shortly after seeding. 476 Notably, channel C1 exhibited significantly different 477 behaviour compared to the other channels (C2–C6), 478 with over 0.56% fewer particles exported and 0.67% 479 more particles stranded.

480 For the artificial channels, the percentage of 481 exported particles increased from 2.3% at C1 to 482 10.1% at C6. Conversely, the total percentage of 483 stranded particles followed an opposite trend, with 484 the highest stranding occurring at C1 (7.85%) and 485 the lowest at C6 (0.053%). This indicates a reduc- 486 tion in beach stranding as the channels transition 487 from Santos toward Guarujá.

488 Additionally, in 2019, particles released at INF1 489 exhibited a significantly higher export rate, with 490 nearly total stranding. The total affected area also 491 decreased along the channels, from 34.062 km<sup>2</sup> at 492 C1 to the smallest area of 17.121 km<sup>2</sup> at C6.

493 In 2020, similar to 2019, the total export in- 494 creased progressively from C1 to C6, although the 495 increment between C5 and C6 was smaller, only 496 0.13%. The total stranding also decreased progres- 497 sively from C1 to C6, but with a smaller decline 498 between the channels compared to 2019. This sug- 499 gests that, as with exports, stranding was also lower 500 in 2020, allowing particles to remain active in the 501 system for a longer period.

502 The total dispersion areas were larger in 2020 503 than in 2019 due to the earlier opening of the flood-

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504 gates—on the second day of 2020 compared to the  
 505 second week of 2019. This earlier opening allowed  
 506 more time for particle dispersion and exposed them  
 507 to stronger currents, increasing the dispersion po-  
 508 tential. However, unlike in 2019, the difference in  
 509 export and stranding between C1 and the other ar-  
 510 tificial channels was much smaller in 2020. The  
 511 difference between C1 and C2 was less than 100  
 512 stranded particles, as well as for those exported.

513 In summary, 17.8% of the particles released in  
 514 2019 were exported, compared to 28.12% in 2020.  
 515 Meanwhile, the proportion of stranded particles de-  
 516 creased from 14.8% in 2019 to 3.7% in 2020 (not  
 517 shown). These values further highlight that the  
 518 exportation suppresses the stranding, and also indi-  
 519 cate that most of the particles (67.4% in 2019 and  
 520 68.18% in 2020) keep active in the water column.

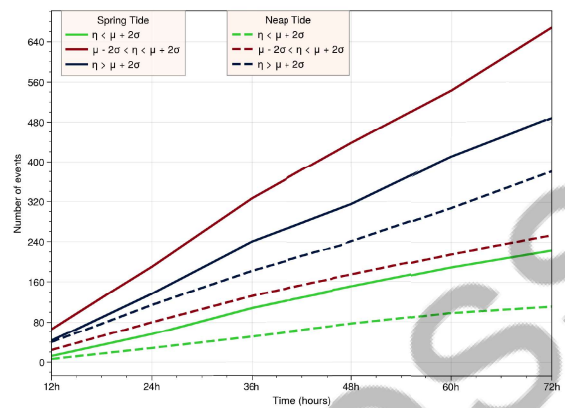
#### 521 TIDAL DYNAMICS INFLUENCE ON MICROPLASTIC 522 RETENTION

523 To understand the microplastic accumulation under  
 524 different elevation scenarios, we computed the prob-  
 525 ability (%) of stranding. Only probabilities higher  
 526 than 5% within all 72-hour periods were used, com-  
 527 puting the total particles considering all seeding  
 528 points. Therefore, this analysis gives an under-  
 529 standing of which sea surface elevation conditions  
 530 the SES accumulates more marine litter.

531 Figure 5 presents the number of events in which  
 532 the probability (%) of contact exceeded 5% over  
 533 time within all 72-hour periods, considering the total  
 534 amount of particles from all launch points. These re-  
 535 sults demonstrate that the system retains microplas-  
 536 tics and increases contamination along the coast.

537 Under extreme subsidence conditions of sea sur-  
 538 face elevation, the number of contact events with  
 539 the shore was the lowest, both in terms of the num-  
 540 ber of affected points and the number of particles,  
 541 regardless of the tidal phase. Thus, lower extreme  
 542 elevations not only reduce particle accumulation  
 543 along the coast but also enhance exportation and  
 544 prolong the retention of microplastics within the wa-  
 545 ter column of the bay.

546 During neap tide, a higher sea surface elevation  
 547 led to an increased number of contacts with the  
 548 mainland and more events where the probability  
 549 of contact exceeded 5%. During spring tides, the



550 **Figure 5.** Number of events with a probability greater  
 551 than 5% of contact with a dry cell over the time in the  
 552 averaged 72h periods for 2019 and 2020 for all particles  
 553 released in INF1, INF2, MID and the artificial channels  
 554 (C1 - C6) according to sea surface height condition ( $\eta$ ),  
 555 mean ( $\mu$ ) and standard deviation ( $\sigma$ ).

556 greatest number of events with a contact probabili-  
 557 ty above 5%, as well as a higher concentration of  
 558 contaminated points (particularly in the interior of  
 559 the Santos Channel and the São Vicente mouth),  
 560 occurred under normal sea surface elevation condi-  
 561 tions. Additionally, for all elevation conditions, the  
 562 total number of contact events exceeding 5% prob-  
 563 ability was higher during spring tides.

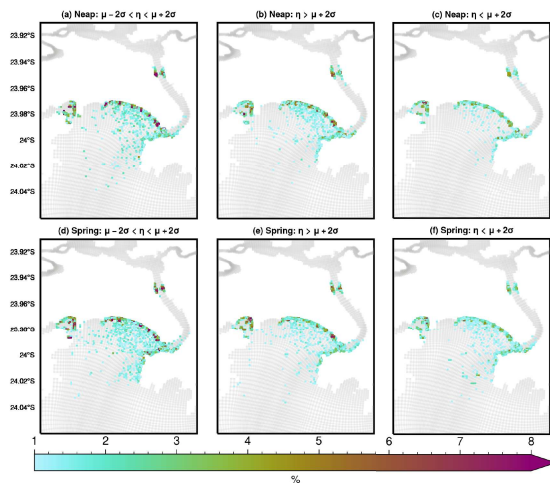
564 Thus, during spring tides, the highest contami-  
 565 nation probability occurs under normal sea surface  
 566 elevation conditions, whereas during neap tides,  
 567 it is associated with extreme elevations. However,  
 568 extreme subsidence consistently resulted in the low-  
 569 est contamination levels across all surface elevation  
 570 conditions.

571 To assess the likelihood of contamination, the  
 572 three-dimensional probabilities of contact after 72  
 573 hours of simulation are presented for all seeding  
 574 points with potential contamination in Santos Bay.  
 575 These probabilities are expressed as percentages  
 576 for each sea surface elevation condition. Addition-  
 577 ally, the regions associated with a contact probabili-  
 578 ty greater than 5% are highlighted.

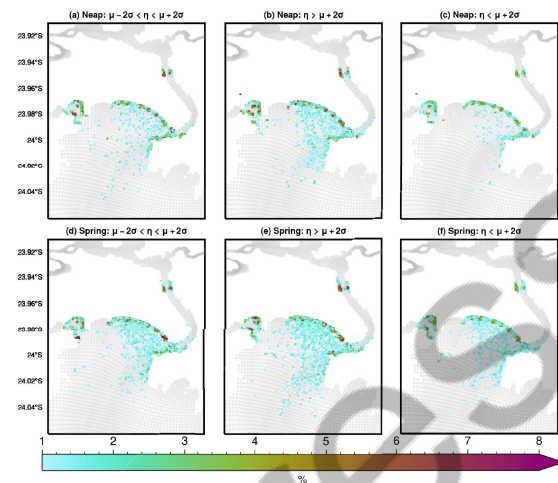
579 Figure 6 illustrates these probabilities for neap  
 580 and spring tides. Similarly, Figure 7 presents the  
 581 corresponding analyses for the year 2020.

582 In 2020, particle dispersion extended to the out-  
 583 ermost regions of the Southeastern Shelf (SES),

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**Figure 6.** Probability of three-dimensional contactin SES during neap tide (a,b,c) and spring tide (d,e,f), after 72h of particles released in MID, INF1, IN2 and in channels C1 to C6 in 2019 between 0 and -5 m for  $\mu - 2\sigma < \eta < \mu + 2\sigma$  (a,d),  $\eta > \mu + 2\sigma$  (b,e), and  $\eta < \mu + 2\sigma$  (d,f). According to sea surface height condition ( $\eta$ ), mean ( $\mu$ ) and standard deviation ( $\sigma$ ).



**Figure 7.** Probability of three-dimensional contact in SES during neap tide (a,b,c) and spring tide (d,e,f), after 72h of particles released in MID, INF1, IN2 and in channels C1 to C6 in 2020 between 0 and -5 m for  $\mu - 2\sigma < \eta < \mu + 2\sigma$  (a,d),  $\eta > \mu + 2\sigma$  (b,e), and  $\eta < \mu + 2\sigma$  (d,f). According to sea surface elevation condition ( $\eta$ ), mean ( $\mu$ ) and standard deviation ( $\sigma$ ).

578 reaching areas near Guarujá. However, no contact 603  
 579 probabilities exceeding 5% were observed beyond 604  
 580 the channel exit between Santos and Guarujá or 605  
 581 within Santos Bay (Figure 6). 606

582 Apart from the regions influenced by particles released 607  
 583 at the MID point and the channel entrances near INF1 608  
 584 and INF2, no significant contact probability was detected 609  
 585 within the SES or in the direction of Bertioga for any of 610  
 586 the years studied. 611

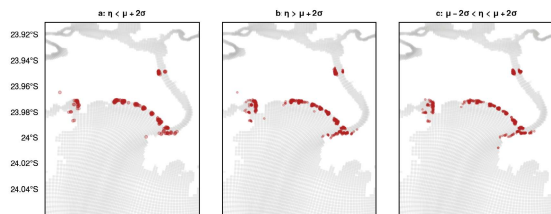
587 For the 72-hour periods, the largest accumulations 612  
 588 of particles, in both 2019 and 2020, during neap and 613  
 589 spring tides – occurred along the coast (Figures 6 and 7). 614  
 590 Contact probabilities greater than 5% were observed only 615  
 591 on the beaches of Santos Bay, at the mouths of the Santos 616  
 592 and São Vicente channels, and near the seeding points in the 617  
 593 middle estuary. Contact events increased during each 12-  
 594 hour period, with the lines representing the number of  
 595 events with a contact probability greater than 5%  
 596 becoming steeper toward the end of the period 5.  
 597

598 The lowest number of hotspots is observed under 600  
 599 extreme subsidence conditions (Figures 9 a and 9 b).  
 601 The highest number of hotspots occurs under extreme  
 602 elevation conditions (Figures 8 b and 9 b), although it is only slightly higher than under normal

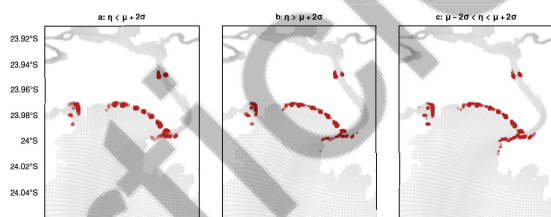
elevation conditions during neap tide (Figure 8 c).

A greater number of hotspots is observed with the highest elevations, particularly in the middle portion of the channel between Santos and Guarujá, with an increase extending toward Guarujá and into the channel between Guarujá and Santos. Between the artificial channel C1 and the mouth of the São Vicente channel, no accumulation hotspots were detected. However, in the latter, there is a noticeable increase in the number of points more susceptible to microplastic contamination as surface elevation rises. The entire shore of Santos Bay, between the artificial channels, presents accumulation hotspots for microplastics under any sea surface elevation condition ( $\eta$ ) (Figures 8 and 9).

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**Figure 8.** Hotspots (location with contact probability higher than 5 %) of microplastic accumulation in the SES during neap tide for  $\eta < \mu + 2\sigma$  (a),  $\eta > \mu + 2\sigma$  (b) and  $\mu - 2\sigma < \eta < \mu + 2\sigma$  (c).



**Figure 9.** Hotspots (location with contact probability higher than 5 %) of microplastic accumulation in the SES during spring tide for  $\eta < \mu + 2\sigma$  (a),  $\eta > \mu + 2\sigma$  (b) and  $\mu - 2\sigma < \eta < \mu + 2\sigma$  (c).

## DISCUSSION

The largest dispersion in Santos Bay occurs when particles are seeded at the mouth of the São Vicente channel (Figures 6 and 7). These particles have the potential to reach the entire Bay of Santos and São Vicente, but with limited inland dispersion. This dispersion is primarily influenced by the tidal cycle, with exportation during ebb tide and accumulation during flood tide (Malli et al. 2022).

C6 is the closest station to the Porto Channel and, consequently, to the INF1 release point. This proximity resulted in contrasting outcomes between 2019 and 2020, with greater particle accumulation observed in 2019 and higher export in 2020. However, due to the dispersion patterns and the combined effects of particle export and stranding from INF2 and the artificial channels, the overall export of particles from the mouth of the Porto Channel to the South Brazil Bight (SBB) was more pronounced.

The preferential southwestward flow in the inner region of the SBB under high-pressure atmospheric synoptic conditions (Dottori & Castro 2009) enhances the export of microplastics from sources along this direction. This effect is further amplified by increased river discharge associated with the Porto channel, driven by deeper waters due to dredging activities and the connection between the Piaçaguera and Bertioga channels (Figure 1) (Pinheiro et al. 2021). Additionally, surface Stokes drift dynamics linked to the wave climate (Stein & Siegle 2019, Dagestad & Röhrs 2019), cold front incursions (Stech & Lorenzetti 1992), and buoyancy advection from fresher waters toward the Equator contribute to the annual-scale dispersion of particles toward Guarujá and Bertioga. This combination of forces is also thought to be responsible for the drift of the exported particles from SES to the SBB towards the Equator, until they meet more complex dynamics. However, further study is needed for that conclusion.

The weakening of the tidal amplitude relative to river flow during neap tides intensifies ebb currents (Seiler et al. 2020). As a result, particles released at INF1 and within the artificial channels near the São Vicente channel exhibit a greater potential for dispersion and export, driven by stronger ebb currents at these locations compared to the Harbour

channel (Figure 5).

The exportation results show that the SES has a greater annual potential for exporting microplastics to SBB than retention, while the hydrodynamic and differences associated with the Lagrangian model (Dagestad & Röhrs 2019) that made 2020 export more. The probability of contact demonstrated that in the SES, there is a greater accumulation of microplastics on the coast with increasing sea surface elevation, corroborating other studies (Gorman et al. 2020, Vermeiren et al. 2016, da Silva Ferreira et al. 2021).

The spring tide redistributes free particles in the water column that have remained since the neap tide (Malli et al. 2022). That is, during neap tide, the greater the elevation of the free surface, the greater the accumulation of particles on the shore. During the spring tide, however, the highest flood and ebb velocities redistribute the particles that remained active in the water column (Figure 5). During spring tide, favourable hydrodynamic conditions accumulate the remaining particles, from the previous neap tides on the coast, both under normal and extreme elevations.

Our results indicate an increase in contact points along the coast, leading to intensified beaching of particles near the port channel, particularly during spring tides in both 2019 and 2020. This pattern is primarily due to the lower ebb tide velocity compared to the flood tide in this region (Seiler et al. 2020).

The combination of ebb and flood tide variations, the limited area available for dispersion (Piehl et al. 2020), and the considerable depth relative to the narrow width of the channel forces surface and subsurface particles to run aground before they can reach lower-energy zones and disperse over a longer period (Röhrs et al. 2014). As a result, a higher frequency of events with contact probabilities exceeding 5% occurred during normal elevation phases of the spring tide.

In summary, the flood tide during spring promotes particle accumulation, while the regional ebb tide under normal elevation conditions is insufficient to remove them, leading to prolonged retention. However, during extreme elevations in spring tide, when astronomical forcing surpasses meteorological influences, the ebb tide becomes strong

713 enough to resuspend particles into the water col- 759  
714 umn, preventing prolonged contact with the coast. 760

715 Regarding accumulation and the areas most sus- 761  
716 ceptible to microplastic pollution in the Santos Estu- 762  
717 arine System (SES), our results indicate that over 763  
718 short periods, comparable to a few tidal cycles, par- 764  
719 ticles tend to accumulate and run aground near their 765  
720 release points, regardless of whether it is a spring 766  
721 or neap tide. 767

722 However, in annual simulations, accumulation be- 768  
723 comes more extensive, spreading along the entire 769  
724 coast of Santos and São Vicente Bay, although con- 770  
725 centrations remain higher near the sources. This 771  
726 pattern suggests that, similar to what was observed 772  
727 in Praia da Enseada do Guarujá (da Silva Ferreira 773  
728 et al. 2021), plastic accumulation in the SES is influ- 774  
729 enced by both proximity to the source and the mor- 775  
730 phological characteristics of the region. More im- 776  
731 portantly, the frequency of extreme elevation events 777  
732 plays a crucial role in determining the overall distri- 778  
733 bution and retention of microplastics. 779

## 734 SUMMARY 781

735 This study brings to SES the coupling between an 783  
736 Eulerian model, a spectral wave model and a La- 784  
737 grangian model in offline mode. Such an approach 785  
738 is increasingly applied to studies of world ocean 786  
739 pollution, optimizing simulation time and error prop- 787  
740 agation of hydrodynamic fields drifting from floating 788  
741 microplastics neutral to quantify the potential for 789  
742 export, accumulation and dispersion of this growing 790  
743 form of pollution, identifying the regions most sus- 791  
744 ceptible to contamination under different conditions 792  
745 of elevation of the sea surface. 793

746 Annually, through the density gradients and 794  
747 quantification of particles that leave the numerical 795  
748 grid of the SES and those that become attached 796  
749 to the dry cells, the SES has a greater potential 797  
750 for exporting microplastics to SBB (23% of the total 798  
751 released) than accumulation (8.9% of the total 799  
752 released). Most of the microplastics released annu- 800  
753 ally in the SES (68.1%), however, remain active in 801  
754 the water column of the middle estuary and Santos 802  
755 Bay. 803

756 The hydrodynamic configuration of the system, 804  
757 characterised by a preferential southwestward flow, 805  
758 Stokes drift pushing waves toward the coast, hy- 806

drodynamic inversions due to cold front intrusions,  
and buoyancy-driven advection toward the north-  
east, collectively influence the movement of neu-  
trally buoyant particles. These processes maintain  
particle activity within the Bay of Santos and the  
channels of Porto and São Vicente.

Thus, the SES exhibits a hydrodynamic regime  
that favours the sustained suspension of seeded  
particles in the water column, preventing their rapid  
deposition and prolonging their residence time in  
the system.

Extreme sea surface subelevation, defined as  
values below two standard deviations from the  
mean, reduces particle accumulation along the  
coast and enhances dispersion throughout Santos  
Bay, regardless of whether it occurs during spring  
or neap tides.

Conversely, sea surface elevations exceeding  
two standard deviations from the mean promote  
greater particle accumulation along the coast and  
reduce dispersion within Santos Bay. This effect  
is particularly pronounced toward the south due  
to hydrodynamic inversions associated with such  
extreme elevation conditions.

The spring tide plays a crucial role in transporting  
particles toward the coast that were not advected  
to the South Brazil Bight (SBB) during the neap  
tide. This transport pathway highlights the vulner-  
ability of the Santos Estuarine System (SES) to micro-  
plastic contamination, particularly under extreme  
sea surface elevation events such as storm surges.

The areas most susceptible to microplastic con-  
tamination are those closest to the primary sources  
of microplastics, as well as regions influenced by  
channel curvature, width, and ebb tide velocity,  
which determines the efficiency of particle export  
from the estuary.

Despite variations in sea surface elevation, the  
SES remains an effective exporter of microplastics  
to the SBB. However, within the SES itself, hydro-  
dynamic processes were not sufficient, even in annual  
simulations, to transport microplastics to locations  
further north of Guarujá. A three-dimensional anal-  
ysis of contact probabilities further indicated that  
microplastic particles did not significantly enter the  
estuarine channels, as they tended to quickly run  
aground along the channel sides or settle on the  
bottom.

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807 Although this study provides valuable insights 845  
 808 into microplastic transport within the SES, further 846  
 809 research is needed to refine our understanding of 847  
 810 their dispersion patterns. Future studies should 848  
 811 investigate microplastic behaviour under various  
 812 extreme conditions, such as atmospheric blocking 849  
 813 events, which induce sea surface subelevation while 850  
 814 reducing river discharge (Silva & Dottori 2021), or 851  
 815 extreme precipitation scenarios. Additionally, simu- 852  
 816 lations with expanded numerical domains are nec- 853  
 817 essary to track the fate of microplastics after they 854  
 818 reach the SBB, which was beyond the scope of this  
 819 study.

## 820 AUTHOR CONTRIBUTION

821 **A.D.** : Conceptualisation; Formal analysis; Investi- 845  
 822 gation; Writing – original draft; Writing – review 846  
 823 & editing; 847

824 **D.A.S.** : Conceptualisation; Investigation; Writing – 848  
 825 review, 849

826 **D.K.S.** : Investigation; Writing, 850

827 **M.D.** : Conceptualisation, Supervision; Resources; 851  
 828 Project Administration; Funding Acquisition; 852  
 829 Writing - Review. 853

## 830 DATA AVAILABILITY STATEMENT

831 The codes for both Lagrangian and Eulerian analy- 845  
 832 sis are available by contacting the authors. 846

## 833 CONFLICTS OF INTEREST

834 The authors declare that they have no conflict of 845  
 835 interest. 846

## 836 FUNDING

837 The authors would like to acknowledge and thank 845  
 838 Petróleo Brasileiro S.A. (PETROBRAS) for their fi- 846  
 839 nancial support, provided through the USP Foun- 847  
 840 dation (FUSP), for the duration of this project from 848  
 841 2019 to 2022. 849

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